

## Glucose-induced nitrate assimilation in prairie and cultivated soils

T. H. DeLUCA<sup>1,\*</sup> & D. R. KEENEY<sup>2</sup>

<sup>1</sup> Free Department, Slippery Rock University, Slippery Rock, PA 16057, USA (\*Corresponding author); <sup>2</sup> Leopold Center for Sustainable Agriculture, 126 NSTL, Iowa State University, Ames, Iowa 50010

Received 29 March 1993; accepted 27 May 1993

**Key words:** amino acids, assimilatory nitrate reductase, immobilization, N cycling

**Abstract.** Native prairie and grassland soils are known to accumulate little inorganic N; however,  $\text{NO}_3^-$  is constantly being formed and re-immobilized. This suggests that microorganisms in prairie soils would be highly efficient in the assimilation of  $\text{NO}_3^-$  and would regularly have the assimilatory  $\text{NO}_3^-$  reductase (ANR) enzyme in an induced and active state. Aerated slurries and static systems prepared from prairie and cultivated soils amended with glucose and  $\text{NO}_3^-$  were observed for changes in  $\text{NO}_3^-$  concentration with time. Nitrate assimilation in the presence of glucose occurred more rapidly in cultivated than in prairie soils from the same soil map unit. Nitrate assimilation rates were not affected by inoculation of prairie soil with cultivated soil. It has been reported that the addition of glucose and  $\text{NO}_3^-$  to soils results in increased peptidase activity and a release of free amino acids. Mixing, sieving, and slurring of prairie soils followed by treatment with glucose and  $\text{NO}_3^-$  may release free amino acids and other ANR inhibitors into the prairie soil slurries. Prairie soils had higher concentrations of soluble amino-N than cultivated soils with or without glucose and  $\text{NO}_3^-$  additions. Prairie soils also had greater concentrations of total Kjeldahl N and readily hydrolyzed amino acids than corresponding cultivated soils.

## Introduction

It is well established that native or established grassland soils accumulate little inorganic N, although  $\text{NH}_4^+$  and  $\text{NO}_3^-$  are continually being formed (Woodmansee et al. 1981; Schimel et al. 1989). Nitrification has been found to account for about one third of total mineralization in a grassland soil where there is little or no  $\text{NO}_3^-$  accumulation (Schimel et al. 1989). This relatively high activity of nitrifying bacteria in grassland soils has been confirmed in other studies (Davidson et al. 1990; Both et al. 1992). The lack of  $\text{NO}_3^-$  accumulation greatly reduces the potential for both  $\text{NO}_3^-$  leaching and gaseous losses of N (Woodmansee et al. 1981; Goodroad & Keeney 1984).

It has been suggested that the lack of  $\text{NO}_3^-$  accumulation in prairie or grassland soils is a function of plant uptake (Clark 1977) and microbial immobilization of  $\text{NO}_3^-$  (Schimel et al. 1989; Davidson et al. 1990). Microbes have been shown to out-compete plants for  $\text{NO}_3^-$  in aerobic short-term studies with grassland soils (Jackson et al. 1989). It is likely that microorganisms are assimilating  $\text{NO}_3^-$  because there is limited denitrification in prairie soils (Goodroad & Keeney 1984).

Studies have clearly shown that soil microorganisms preferentially assimilate  $\text{NH}_4^+$  over  $\text{NO}_3^-$  and assimilate little  $\text{NO}_3^-$  in the presence of  $\text{NH}_4^+$  or several amino acids (Rice & Tiedje 1989; McCarty & Bremner 1992). Rice & Tiedje (1989) present a critical level of  $0.1 \mu\text{g NH}_4^+\text{-N g}^{-1}$  soil, below which  $\text{NO}_3^-$  is readily assimilated. Ammonium is generally found at greater levels than  $\text{NO}_3^-$  in grassland soils (Woodmansee et al. 1981). It was thus suggested that  $\text{NO}_3^-$  assimilation in prairie soils occurred in microsites where  $\text{NH}_4^+$  had been exhausted by microbial activity (Davidson et al. 1990).

Laboratory studies using disturbed grassland and cultivated soils have demonstrated higher rates of gross  $\text{NO}_3^-$  immobilization in cultivated soils than in grassland soils (Schimel 1986). Greater N mineralization in disturbed prairie soils compared to cultivated soils (Keeney & Bremner 1964) may eliminate the potential for measuring gross immobilization *in-vitro* without an added energy source. To correct for this, Rice & Tiedje (1989) added glucose to soils to induce measurable  $\text{NO}_3^-$  assimilation.

Microbial assimilation of  $\text{NO}_3^-$  requires the presence of sufficient energy as C to reduce  $\text{NO}_3^-$  to  $\text{NH}_4^+$  before incorporation into amino acids. Prairie soils have been found to have higher concentrations of labile organic C and N than their cultivated counterparts (Schimel 1986; Woods 1989; DeLuca & Keeney 1993). This pool of labile C would supply the necessary energy to account for the high rate of  $\text{NO}_3^-$  assimilation observed in prairie soils.

It is not clear whether  $\text{NO}_3^-$  assimilation in the presence of added glucose would occur more rapidly in prairie or cultivated soils based on existing studies. It would appear that assimilatory  $\text{NO}_3^-$  reductase (ANR) activity would be high in  $\text{NH}_4^+$  depleted microsites of prairie or grassland soils. Cultivated soils that accumulate  $\text{NO}_3^-$  and lack sufficient C to allow for microbial assimilation of  $\text{NO}_3^-$  would not normally have the  $\text{NO}_3^-$  reductase enzyme induced and should thus be less efficient at reducing  $\text{NO}_3^-$  than their prairie counterparts.

The objectives of this study were to observe the rate of glucose-induced assimilatory  $\text{NO}_3^-$  reduction in prairie and cultivated soils from the same soil map unit and to determine possible constraints on  $\text{NO}_3^-$  reduction beyond C limitation or inhibitory concentrations of  $\text{NH}_4^+$ .

## Materials and methods

Soils were collected from paired prairie and cultivated sites in central Iowa previously described by DeLuca & Keeney (1993). All soils were sieved through a 2-mm mesh and stored at 5 °C until time of use (Table 1).

Table 1. General soil properties.

| Soil                  | Cover   | Sand                          | Silt | Clay | Total carbon | pH  |
|-----------------------|---------|-------------------------------|------|------|--------------|-----|
|                       |         | -----g kg <sup>-1</sup> ----- |      |      |              |     |
| Kossuth <sup>1</sup>  | Crop    | 180                           | 300  | 520  | 32.6         | 6.1 |
|                       | Prairie | 250                           | 330  | 420  | 39.0         | 5.9 |
| Clarion <sup>2</sup>  | Crop    | 320                           | 460  | 220  | 29.5         | 5.9 |
|                       | Prairie | 370                           | 460  | 170  | 40.5         | 5.9 |
| Canisteo <sup>3</sup> | Crop    | 270                           | 370  | 360  | 28.7         | 7.3 |
|                       | Prairie | 250                           | 350  | 400  | 40.5         | 7.1 |

<sup>1</sup> Kossuth = Typic Haplaquolls

<sup>2</sup> Clarion = Typic Hapludolls

<sup>3</sup> Canisteo = Typic Haplaquolls (calcareous)

Assimilatory NO<sub>3</sub><sup>-</sup> reduction studies in aerated soil slurry were performed as described by McCarty & Bremner (1992). To induce the NO<sub>3</sub><sup>-</sup> reductase enzyme and assimilate pre-existing NH<sub>4</sub><sup>+</sup> preceding initiation of the study, field-moist 10 g (oven dried equivalent) soil samples were treated with 25 mg of C as glucose and 1.8 mg N as KNO<sub>3</sub>, brought to 60% water-holding capacity, and then incubated at 30 °C for 16 hours. Soils were then treated with 5.0 mg C as glucose and 0.6 mg N as KNO<sub>3</sub> and shaken with 30 ml of distilled deionized water to obtain the soil slurries reported in this study.

The aerated soil slurries were continuously monitored for levels of NO<sub>3</sub><sup>-</sup> by using a NO<sub>3</sub><sup>-</sup>-sensitive electrode. The soil slurries were treated with 14 mg of K<sub>2</sub>SO<sub>4</sub> (to adjust ionic strength), placed on magnetic stir plates, maintained at 30 °C in a constant temperature chamber, and aerated by bubbling with a stream of air. A NO<sub>3</sub><sup>-</sup> electrode (Orion model 93-07) coupled with a double-junction reference electrode (Orion model 90-02) containing 0.4 M K<sub>2</sub>SO<sub>4</sub> outer filling solution was placed into each slurry. Electrodes were calibrated against KNO<sub>3</sub> standards prior to each analysis. The millivolt output was then continuously charted with a pen recorder connected to the electrode meter. To determine whether cross-

inoculation of soils would increase or decrease the rate of  $\text{NO}_3^-$  reduction, 9 g (oven dried equivalent) of prairie soil was amended with 1 g of cultivated soil.

To monitor the reduction of  $\text{NO}_3^-$  in soil over a 4-day period, 10 g (oven-dried equivalent) soil samples were placed in 250 ml French square bottles, treated with 25 mg C as glucose and 1.8 mg N as  $\text{KNO}_3$ , brought to 60% water-holding capacity, and allowed to incubate at  $30^\circ\text{C}$  for 4 days. After 0, 1, or 4 days of incubation, 50 ml of 2 M KCl were added to each bottle, and the bottles were shaken for 1 hr. Suspensions were then filtered through Whatman number 5 filter papers and analyzed for  $\text{NH}_4^+$  and  $\text{NO}_3^-$  with a flow injection auto analyzer (Lachat QuickChem Method No. 12-107-06-2-A and 12-107-0401B, Milwaukee, WI).

Soluble amino-N was determined in soils following pre-treatment of soils with glucose and  $\text{KNO}_3$  as described above or with water. Soil samples (10 g oven dried equivalent) were wetted to 60% water holding capacity with either distilled water or with water plus 25 mg of C as glucose and 1.8 mg N as  $\text{KNO}_3$ . After 16 hours the water treated soils were amended with 30 ml of distilled water and the glucose treated soils were amended with 30 ml of distilled deionized water containing 5.0 mg C as glucose and 0.6 mg N as  $\text{KNO}_3$ . Samples were then shaken for 30 minutes and then the reaction stopped by adding 30 ml of 4 M KCl to all samples (to obtain a 2 M KCl extract). The soil slurries were then shaken for 5 minutes and filtered through  $1.2\ \mu\text{m}$  glass fiber filter paper and analyzed for amino-N as described in DeLuca & Keeney (1993).

An assay of amino acids that are readily hydrolyzed was determined by placing 3 g of 100-mesh sieved oven-dried soil in a long Kjeldahl tube and adding 2 drops of octyl alcohol and 20 ml 1 N HCl. The tubes were then heated to  $125^\circ\text{C}$  for 1 hr. The soil hydrolysate was filtered through Whatman number 42 filter paper, and the pH of the filtrate was slowly adjusted to 5.0 by using 1 N NaOH. The filtrate was then quantitatively transferred to a 100 ml volumetric flask and brought to volume. The hydrolysate was then analyzed for total amino acids (Stevenson 1982).

Total Kjeldahl N was determined by using a semimicro Kjeldahl method (Bremner & Breitenbeck 1983). Total C in soils was determined by the Walkley-Black method (Nelson & Sommers 1982). Particle size distribution was determined by hydrometer (Gee & Bauder 1986).

## Results and discussion

Nitrate reduction occurred readily in the cultivated soils, but appeared greatly inhibited in the prairie soils (Fig. 1). The average rate of  $\text{NO}_3^-$

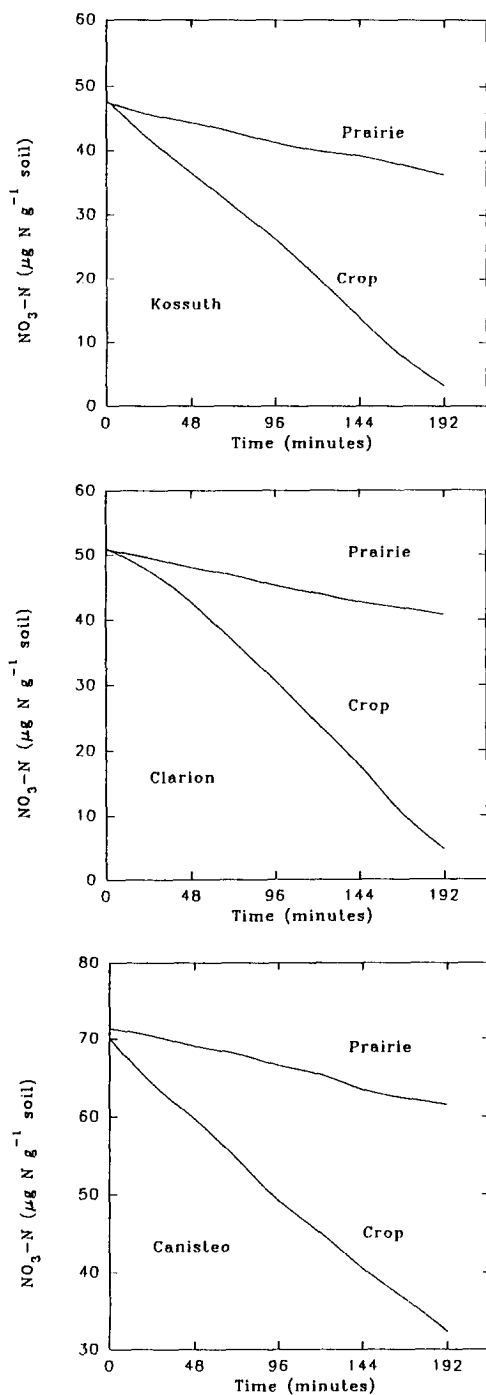


Fig. 1. Change in  $\text{NO}_3^-$  concentration with time in Kossuth, Clarion, and Canisteo prairie and cultivated soil slurries treated with glucose and  $\text{KNO}_3$  and incubated at  $30^\circ\text{C}$ .

assimilation across all three soil types was  $13.8 (+/-1.8) \mu\text{g NO}_3^- \text{ hr}^{-1}$  in cultivated soils and  $3.24 (+/-0.12) \mu\text{g NO}_3^- \text{ hr}^{-1}$  in prairie soils. The low level of assimilatory  $\text{NO}_3^-$  reductase (ANR) activity in the prairie soils continued throughout the experiments with no observed increase in the rate of  $\text{NO}_3^-$  reduction. Ammonium was generally at undetectable levels in all soils at the end of the slurry experiments, and addition of  $\text{NH}_4^+$  to soil slurries rapidly inhibited  $\text{NO}_3^-$  reduction, demonstrating that the change in  $\text{NO}_3^-$  concentration with time was due to assimilation rather than denitrification of the applied  $\text{NO}_3^-$ .

The 16 hr preincubation of soils with glucose and  $\text{KNO}_3$  generally resulted in the complete reduction of  $\text{NO}_3^-$  present in the cultivated soils, whereas much of the original  $\text{NO}_3^-$  applied to the prairie soil remained at the end of 16 hr. This suggests that the lack of  $\text{NO}_3^-$  reduction in prairie soils in the soil slurry study was not due to a missed lag phase of bacterial growth on glucose and  $\text{NO}_3^-$ .

Four-day, non-slurry incubations of the prairie and cultivated soils with added glucose and  $\text{NO}_3^-$  resulted in the immobilization of most of the applied  $\text{KNO}_3$  after 1 day of incubation in the cultivated soils and after 4 days of incubation in prairie soils (Table 2). This demonstrates that, even in the presence of limited soil microsites, ANR activity (induced by glucose) occurs more slowly in prairie soils than in cultivated counterparts.

Inoculation of prairie soil with cultivated soil did not significantly affect the rate of  $\text{NO}_3^-$  assimilation in the prairie soils (data not shown). It is likely, therefore, that the inhibition of  $\text{NO}_3^-$  assimilation in prairie soils is not due to a lack of ANR active organisms, but rather to a physical or chemical inhibition of ANR in the prairie soil.

Nitrifiers have been shown to be poor competitors for  $\text{NH}_4^+$  in energy limited chemostats and in soil columns (Verhagen & Laanbroek 1991; Verhagen et al. 1992). It is thus unlikely that the low rate of  $\text{NO}_3^-$  assimilation in the prairie soils is a function of a high rate of nitrification in the prairie soil slurries as there would be no C-limited microsites that would give the nitrifiers a competitive advantage.

Extensive nitrification or immobilization of  $\text{NH}_4^+$  results in the release of fixed  $\text{NH}_4^+$  from clay lattices (Drury & Beauchamp 1991). A release of fixed  $\text{NH}_4^+$  after glucose-induced immobilization of all exchangeable  $\text{NH}_4^+$  might also inhibit  $\text{NO}_3^-$  assimilation. However,  $\text{NH}_4^+$  levels in prairie soils treated with glucose and  $\text{KNO}_3$  were not significantly different from  $\text{NH}_4^+$  levels in cultivated soils (data not shown).

Peptidase activity in soil has been found to increase immediately after treatment of soils with glucose and  $\text{NO}_3^-$ , and after 32 hr, there is an increase in nonspecific proteinase activity (Ladd & Paul 1973). Treatment

Table 2. Nitrate concentrations in three prairie and cultivated soils untreated or treated with glucose and  $\text{KNO}_3$  and incubated at 30 °C for 0, 1, and 4 days.

| Soil/<br>Time<br>(day)                  | Cultivated             |         | Prairie |         |
|---|------------------------|---------|---------|---------|
|   | Control                | Treated | Control | Treated |
|   | $\mu \text{ g g}^{-1}$ |         |         |         |
| Kossuth                                 |                        |         |         |         |
| 0                                       | 17.0                   | 69.0    | 3.8     | 45.0    |
| 1                                       | 18.0                   | 1.9     | 4.6     | 12.8    |
| 4                                       | 18.6                   | 1.4     | 5.8     | 1.2     |
| LSD ( $P < 0.05$ ) within columns = 2.9 |                        |         |         |         |
| Clarion                                 |                        |         |         |         |
| 0                                       | 21.3                   | 56.0    | 12.3    | 52.5    |
| 1                                       | 20.1                   | 0.8     | 14.5    | 17.5    |
| 4                                       | 18.3                   | 1.2     | 18.5    | 1.5     |
| LSD ( $P < 0.05$ ) within columns = 1.4 |                        |         |         |         |
| Canisteo                                |                        |         |         |         |
| 0                                       | 17.7                   | 69.0    | 25.0    | 82.5    |
| 1                                       | 19.3                   | 0.7     | 27.6    | 29.2    |
| 4                                       | 20.4                   | 1.1     | 31.2    | 15.7    |
| LSD ( $P < 0.05$ ) within columns = 1.8 |                        |         |         |         |

of soils with glucose and  $\text{NO}_3^-$  results in a release of certain free amino acids (Putnam & Schmidt 1958; Paul & Schmidt 1961). If the addition of glucose and  $\text{KNO}_3$  to soils induced peptidase activity that, in turn, caused a release of amino acids, then the increase in free amino acids or their deamination production ( $\text{NH}_4^+$ ) would inhibit ANR (Rice & Tiedje 1989; McCarty & Bremner 1992).

The pool of soluble amino-N was found to be greater in the prairie soils than in the cultivated soils when incubated either with water or with glucose and  $\text{NO}_3^-$  (Table 3). Glucose and  $\text{NO}_3^-$  treatment appeared to enhance the presence of free amino-N in the cultivated soils, but prairie soils treated with just water had higher levels of soluble amino-N, than the samples treated with glucose and  $\text{NO}_3^-$ . The lack of a glucose and  $\text{NO}_3^-$  induced increase in free amino-N in the prairie soils may be a function of the high ambient levels of labile C in the prairie soils (Woods 1989) which would normally stimulate peptidase activity. Addition of glucose and  $\text{NO}_3^-$  to the prairie soils would increase microbial activity and possibly reduce the level of measurable soluble amino-N due to rapid immobilization of free amino acids in the presence of glucose. The release of free amino-N from cultivated soils following amendment with glucose and  $\text{NO}_3^-$ , may

Table 3. Soluble amino-N following treatment of soils with water or glucose and  $\text{KNO}_3$ , readily hydrolyzed amino acids, and total N in Kossuth, Clarion, and Canisteo prairie and cultivated soils.

| Soil                             | Land use       | Glucose & $\text{KNO}_3$ treated soluble amino-N <sup>1</sup> | Water treated soluble amino-N <sup>1</sup> | Readily hydrolyzed amino acids <sup>2</sup> | Total Kjeldahl N |
|----------------------------------|----------------|---|--|---|------------------|
| ----- $\mu\text{g g}^{-1}$ ----- |                |   |  |   |                  |
| Kossuth                          | C <sup>3</sup> | 0.56  | 0.16                                       | 162.0                                       | 3,033            |
|                                  | P              | 1.27  | 6.67                                       | 208.3                                       | 3,667            |
| T-test                           |                | ** <sup>4</sup>   | **   | **  | **               |
| Clarion                          | C              | 1.21  | 0.59                                       | 169.7                                       | 2,900            |
|                                  | P              | 2.42  | 9.92                                       | 217.7                                       | 3,733            |
| T-test                           |                | **  | **   | **  | **               |
| Canisteo                         | C              | 1.14  | 0.48                                       | 137.0                                       | 2,933            |
|                                  | P              | 1.60  | 3.93                                       | 207.0                                       | 4,367            |
| T-test                           |                | **  | **   | **  | **               |

<sup>1</sup> Soil amino-N soluble in 2 M KCl extracts following 16 hr pre-treatment with glucose and  $\text{KNO}_3$  or water.

<sup>2</sup> Soils hydrolyzed by heating soils to 125 °C in 1 N HCl for 1 hour.

<sup>3</sup> C = cultivated and P = prairie.

<sup>4</sup> \*\* demonstrates significant mean separation at  $P < 0.01$ .

have been due to a lack of available C, thus glucose addition would stimulate peptidase activity above ambient levels.

The prairie soils used in this study had significantly greater concentrations of readily hydrolyzed amino acids and total Kjeldahl N than did cultivated soils (Table 3). Prairie soils clearly have a larger pool of labile N than do corresponding cultivated soils (Keeney & Bremner 1964; Campbell & Souster 1982; Schimel 1985; Schimel 1986; DeLuca & Keeney 1993).

Both cultivated and prairie soils demonstrated ANR activity, but the prairie soils had lower ANR activity than the cultivated soils. These results are similar to laboratory investigations by Schimel (1986) and were possibly a result of a release of free amino acids or other inhibitors when the soils were mixed, sieved, and slurried with glucose and  $\text{NO}_3^-$ . Under natural conditions,  $\text{NO}_3^-$  assimilation in grassland ecosystems occurs in soil microsites void of  $\text{NH}_4^+$  (Davidson et al. 1990). In our experiments, soil microsites were eliminated or greatly reduced (non-slurry experiment), potentially allowing free amino acids or their deamination product



( $\text{NH}_4^+$ ) to interfere with ANR. The prairie soils have the potential to release greater quantities of free amino acids than cultivated soils, but the role of these free amino acids in undisturbed soils needs further investigation.

## Acknowledgements

We thank the Leopold Center for funding this work, Dr. J. M. Bremner for the use of his laboratory, and Carrie Rigdon for laboratory assistance. Journal paper no. J-15291 of the Iowa Agriculture and Home Economics Experiment Station, Ames, IA. Project Number 0181.

## References

- Both GJ, Gerards S & Laanbroek HJ (1992) Temporal and spatial variation in the nitrate oxidizing bacterial community of a grassland soil. *FEMS Microb. Ecol.* 101: 99–112
- Bremner JM & Breitenbeck GA (1983) A simple method for determination of ammonium in semimicro-Kjeldahl analysis of soils and plant materials using a block digester. *Commun. Soil Sci. Plant Anal.* 14: 905–913
- Campbell CA & Souster W (1982) Loss of organic matter and potentially mineralizable nitrogen from Saskatchewan soils due to cropping. *Can. J. Soil Sci.* 62: 651–656
- Clark FE (1977) Internal cycling of  $^{15}\text{N}$  in shortgrass prairie. *Ecology* 58: 1322–1333
- Davidson EA, Stark MJ & Firestone MK (1990) Microbial production and consumption of nitrate in an annual grassland. *Ecology* 71: 1969–1975
- DeLuca TH & Keeney DR (1993) Soluble organics and extractable nitrogen in paired prairie and cultivated soils of Central Iowa. *Soil Sci.* 155: 219–228
- Drury CF & Beauchamp EG (1991) Ammonium fixation, release, nitrification, and immobilization in high- and low-fixing soils. *Soil Sci. Am. J.* 55: 125–129
- Gee GW & Bauder JW (1986) Particle-size analysis. In: Klute A (Ed) *Methods of Soil Analysis*. Part 1 (pp 383–412). American Society of Agronomy, Madison, WI
- Goodroad LL & Keeney DR (1984) Nitrous oxide emission from forest, marsh, and prairie ecosystems. *J. Environ. Qual.* 13: 448–452
- Jackson LE, Schimel JP & Firestone MK (1989) Short-term partitioning of ammonium and nitrate between plants and microbes in an annual grassland. *Soil Biol. Biochem.* 21: 409–415
- Keeney DR & Bremner JM (1964) Effect of cultivation on nitrogen distribution in soils. *Soil Sci. Soc. Am. Proc.* 28: 653–656
- Ladd JN & Paul EA (1973) Changes in enzymic activity and distribution of acid-soluble, amino-acid nitrogen in soil during nitrogen immobilization and mineralization. *Soil Biol. Biochem.* 5: 825–840
- McCarty GW & Bremner JM (1992) Regulation of assimilatory nitrate reductase activity in soil by microbial assimilation of ammonium. *Proc. Natl. Acad. Sci. USA* 89: 453–456
- Nelson DW & Sommers LE (1982) Total carbon, organic carbon, and organic matter. In: Page AL, Miller RH & Keeney DR (Eds) *Methods of Soil Analysis*. Part 2 (pp 539–580). American Society of Agronomy, Madison, WI

- Paul EA & Schmidt EL (1961) Formation of free amino acids in rhizosphere and non-rhizosphere soil. *Soil Sci. Soc. Am. Proc.* 25: 359–362
- Putnam HD & Schmidt EL (1958) Studies on the free amino acid fraction of soils. *Soil Sci.* 87: 22–27
- Rice CW & Tiedje JM (1989) Regulation of nitrate assimilation by ammonium in soils and in isolated soil microorganisms. *Soil Biol. Biochem.* 21: 597–602
- Schimel DS (1986) Carbon and nitrogen turnover in adjacent grassland and cropland ecosystems. *Biogeochemistry* 2: 345–357
- Schimel DS, Coleman DC & Horton KA (1985) Soil organic matter dynamics in paired rangeland and cropland toposequences in North Dakota. *Geoderma* 36: 201–214
- Schimel JP, Jackson LE & Firestone MK (1989) Spatial and temporal effects on plant-microbial competition for inorganic nitrogen in a California annual grassland. *Soil Biol. Biochem.* 21: 1059–1066
- Stevenson FJ (1982) Nitrogen-organic forms. In: Page AL, Miller RH & Keeney DR (Eds) *Methods of Soil Analysis. Part 2* (pp 625–649). American Society Agronomy, Madison, WI
- Verhagen FJM & Laanbroek HJ (1991) Competition for ammonium between nitrifying and heterotrophic bacteria in dual energy limited chemostats. *Appl. Environ. Microbiol.* 57: 3255–3263
- Verhagen FJM, Duyts H & Laanbroek HJ (1992) Competition for ammonium between nitrifying and heterotrophic bacteria in continuously percolating soil columns. *Appl. Environ. Microbiol.* 58: 3303–3311
- Woods LE (1989) Active organic matter distribution in the surface 15 cm of an undisturbed and cultivated soil. *Biol. Fert. Soils* 8: 271–278
- Woodmansee RG, Vallis I & Mott JJ (1981) Grassland nitrogen. In: Clark FE & Rosswall T (Eds) *Terrestrial Nitrogen Cycles* (pp 443–462). *Ecol. Bull.* Vol. 33. Stockholm